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# MEMORANDUM

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To: Chad Benson, John Hall (BLM)  
CC: Scott Fluor, Hollè Waddell, Paul Griffin (BLM)  
From: Michelle Crabb (BLM) WHB Program Population Biologist  
Date: 1/28/2022  
RE: Statistical analysis for 2021 survey of wild burro abundance in the Black Mountain HMA, AZ

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## Summary Table

Survey Area and Dates	Start date	End date	Area names	Area IDs
	11/1/2021	11/5/2021	Black Mountains HMA	AZ0003
Type of Survey	Simultaneous double-observer			
Aviation Details	Pilots: Cody Johnson, John Kelly, El Aero. Helicopters: N226GM, N910BR			
Agency Personnel	Observers: Chad Benson, John Hall, Erik Duarte, Michelle Crabb (BLM), Erin Butler, Larisa Harding, Brandon Foley (AZGFD), Clayton Merrill (NPS) Helicopter Manager: Marcus Escobedo (BLM)			

## Summary Narrative

In November 2021, Bureau of Land Management (BLM) personnel conducted simultaneous double-observer aerial surveys of the wild burro populations in the Black Mountain herd management area (HMA; Figure 1). Surveys were conducted using methods recommended by BLM policy (BLM 2010) and a recent National Academy of Sciences review (NRC 2013) with detailed field methods described in Griffin et al. (2020). These data were analyzed using methods in Ekernas and Lubow (2019) to estimate sighting probabilities for wild burros, with sighting probabilities then used to correct the raw counts for systematic biases (undercounts) that are known to occur in aerial surveys (Lubow and Ransom 2016), and to provide confidence intervals (which are measures of uncertainty) associated with the abundance estimates.

**Table 1.** Estimated abundance (Estimated No. Burros) is for the number of burros in the surveyed areas at the time of survey. 90% confidence intervals are shown in terms of the lower limit (LCL) and upper limit (UCL). The coefficient of variation (CV) is a measure of precision; it is the standard error as a percentage of the estimated abundance. Number of burros seen (No. Burros Seen) leads to the estimated percentage of burros that were present in the surveyed area, but that were not recorded by any observer (Estimated % Missed).

Area	Age Class	Estimated		Std Err	CV	No. Burros Seen	Estimated % Missed	Estimated No. Groups	Estimated Group Size	Foals Per 100 Adults <sup>b</sup>
		No. Burros	LCL <sup>a</sup> UCL							
<b>Black</b>	<b>Total</b>	2,868	2,746 3,001	84.5	2.9%	2,522	12.1%	632	4.5	9.8
<b>Mountain</b>	<b>Foals</b>	256	237 269	10.0	3.9%	227				
<b>HMA</b>	<b>Adults</b>	2,611	2,493 2,739	79.3	3.0%	2,295				

<sup>a</sup> The lower 90% confidence limit is based on bootstrap simulation results or the number of burros seen, whichever is higher.

<sup>b</sup> The estimated ratio of foals to adults reflects what was observed during this November survey. This ratio does not necessarily represent the full cohort of foals for this year, some of which may have died before the survey or were large enough to be included with the adult count. It is possible that some foals may have been born after the survey, in December 2021.

## Abundance Results

The estimated total wild burro abundance within the surveyed area is reported in Table 1. Observers recorded 516 burro groups, of which 486 burro groups had data recorded properly 'on protocol' and that could be used to compute statistical estimates of sighting probability. Of the 516 groups seen, all 516 observations were used to calculate the abundance estimate. Twenty groups (91 burros) were observed on Lake Mead National Recreation Area, these were not separated out in the abundance estimate because of the extensive boundary and lack of natural or man-made barriers to movement between the Black Mountain HMA and the recreation area. Any burro groups that were seen on two separate occasions (double counted), or that were identified as domestic and privately owned, were not used to calculate abundance; however, such groups can be used to parameterize sighting probability if they were recorded on protocol. Coefficient of variation (Table 1) values of less than 10% indicate high precision resulting from high detection probabilities; values between 10-20% indicate medium precision resulting from lower detection probabilities; and values greater than 20% indicate low precision resulting from very low detection probabilities.

Double observer aerial surveys of burros typically contain unmodeled heterogeneity in detection probabilities (discussed below) that cause abundance estimates to be biased too low. Consequently, the abundance estimate presented in Table 1 is likely to be substantially lower than the true number of burros present in the surveyed area.

For reference, a 2017 double observer burro survey and analysis of Sinbad HMA, UT, underestimated burro abundance by approximately 20% compared to tallies of known individuals (Hennig et al., USGS Fort Collins Science Center, manuscript in review). In the absence of better information, adding approximately 20% to the abundance estimate reported in Table 1 (Estimated No. Burros) may lead to more accurate results. However, it is not possible from the available data for the Black Mountain HMA and the analysis presented here to assess the actual additional percentage that should be added.

The mean estimated size of detected burro groups, after correcting for missed groups, was 4.5 burros/group across the surveyed area, with a median of 3.0 burros/group. There were an estimated 9.8 foals per 100 adult burros at the time of these surveys (Table 1). Surveys flown before July are unlikely to include all foals born this year, while surveys flown during or after July would not include foals that were born this year but died before the survey.

## Sighting Probability Results

The combined front observers saw 62.3% of the burro groups (70% of the burros) seen by any observer, whereas the back seat observers saw 74.5% of all burro groups (80.3% of burros) seen (Table 2). At least one observer (front or back) missed 63.2% of burro groups seen by the other. These results demonstrate that simple raw counts do not fully reflect the true abundance without statistical corrections for missed groups, made possible by the double observer method and reported here. Direct counts from aerial surveys underestimate true abundance because some

animals are missed by all observers; this analysis corrects for that bias (Lubow and Ransom 2016). The analysis method used for the surveyed areas was based on simultaneous double-observer data collected during these surveys.

The sample size of observations following protocol was 486 burro groups; these groups were used to parameterize the detection probability model. Survey datasets with sample size less than 20 groups cannot be analyzed using these methods; sample sizes of 20 to 40 groups are considered low and have high risk of containing unmodeled heterogeneity in sighting probability; sample sizes of 41-100 groups are moderate and can estimate effects of many but likely not all potential sightability covariates; and sample sizes >100 groups are large and can account for most sightability covariates.

Unmodeled heterogeneity in detection probability is a systematic problem in double observer aerial surveys of burros, and solving this problem is an area of active research. Burros are difficult to see from the air, and some types of groups are so difficult to see (e.g., groups that are small, standing still, and in heavy tree cover) that they are practically never detected by any observer. When certain types of groups are never seen, their sightability characteristics cannot be described by any set of covariates, and this class of groups disappears from the analysis. Conversely, other types of groups are easy to see (e.g., large groups in open vegetation, close to the helicopter, and running) and every observer sees them nearly every time. The “easy-to-see” types of groups thereby become over-represented in the data. Furthermore, covariate values that sharply reduce detection probability might never be described and thus cannot be modeled. As a result, because detection probabilities can only be estimated from burro groups that were included in the data set, the double observer model tends to over-estimate detection probability for the burro population as a whole. When the detection probability estimate is biased high, the correction factor for how many groups were missed is biased too low. Consequently, unmodeled heterogeneity in detection probability causes double observer analyses to underestimate true burro abundance.

All models used in the double-observer analysis contained an estimated intercept common to all observers. Informed by *a priori* reasoning and preliminary analyses showing overwhelming support, I also included an additional parameter in all models for effects; (1) burro group size; (2) distance of burros from the flight path; (3) observations by front-seat observers on the pilot’s side; (4) effect of pilot CJ, and (5) effect for individual back seat observers. I evaluated 4 possible effects on sighting probability by fitting models for all possible combinations with and without these effects, resulting in 16 alternative models. The 4 effects examined were: (1) high contrast lighting; (2) rugged topography; and (3) percent vegetation cover (squared); and (4) burro group activity. Due to preliminary analyses, I did not consider effects on detection probability of visual field and non-squared percent vegetation due to higher explanatory power of percent vegetation squared. Covariates and their relative effect on sighting probability are shown in Table 3.

Groups that were recorded on the centerline, directly under the aircraft, were not available to backseat observers. For these groups, backseat observers' sighting probability was therefore set to 0. Sighting probability for groups visible on both sides of the aircraft was computed based on the assumption that both backseat observers could have independently seen them, thereby

increasing total detection probability for these groups relative to groups available to only one side of the helicopter.

There was strong support for the effect of rugged terrain (94.9% of AICc model weight). There was moderate support for group activity (46.4%), high contrast lighting condition (32.8%), and percent concealing vegetation (squared, 38.4%). As expected, visibility was higher for burro groups that were larger, closer, moving, in high contrast lighting, in smooth terrain, and lower for groups on the pilot's side, and in greater vegetation (Table 3).

Estimated overall sighting probabilities,  $\hat{p}$ , for the combined observers ranged across burro groups from 0.41-1.00. Sighting probability was  $<0.7$  for 71 (14%), and  $<0.5$  for 11 (2%) of observed groups. Comparing actual burros seen to the estimated abundance computed from the overall  $\hat{p}$ , an estimated 12.1% of the burros present during the survey were never seen by any of the observers (Table 1).

### Assumptions and Caveats

Results from this double observer analysis are a conservative estimate of abundance. True abundance values are likely to be higher, not lower, than abundance estimates in Table 1 because of several potential sources of bias listed below. Results should always be interpreted with a clear understanding of the assumptions and implications.

1. The results obtained from these surveys are estimates of the burros present in the surveyed area at the time of the survey and should not be used to make inferences beyond this context. Abundance values reported here may vary from the annual March 1 population estimates for the HMA; aerial survey data are just one component of all the available information that BLM uses to make March 1 population estimates. Aerial surveys only provide information about the area surveyed at the time of the survey, and do not account for births, deaths, movements, or any management removals that may have taken place afterwards.
2. Simultaneous double-observer analyses cannot account for undocumented animal movement between, within, or outside of the surveyed area. Fences and topographic barriers can provide deterrents to animal movement, but even these barriers may not present continuous, unbroken, or impenetrable barriers. It is possible that the surveys did not extend as far beyond a boundary as burros might move. Consequently, there is the possibility that temporary emigration from the surveyed area may have contributed to some animals that are normally resident having not being present at the time of survey. In principle, if the level of such movement were high, then the number of animals found within the survey area at another time could differ substantially. If there were any wild burros that are part of a local herd but were outside the surveyed areas, then Table 1 underestimates true abundance.
3. The validity of the analysis rests on the assumption that all groups of animals are flown over once during a survey period, and thus have exactly one chance to be counted by the front and back seat observers, or that groups flown over more than once are identified and considered only once in the analysis. Flight line transects were spaced  $\frac{1}{2}$  mile apart, so there was nominally an

opportunity to observe burros throughout the entire surveyed area. The drop-off in detection probability as a function of distance (Table 3) was steep, but even for animals located at more than 400 m from the transect line, there was a non-zero probability of detection. Those lower detections for distance groups translated to higher correction factors for those groups, in their contribution to the estimated total abundance value (Table 1). Animal movements during a survey can potentially bias results if those movements result in unintentional over- or undercounting of burros. Groups counted more than once would constitute ‘double counting,’ which would lead to estimates that are biased higher than the true number of groups present. Groups that were never available to be seen (for example due to temporary emigration out of the study area or undetected movement from an unsurveyed area to an already-surveyed area) can lead to estimates that are negatively biased compared to the true abundance.

Survey SOPs (Griffin et al. 2020) call for observers to identify and record ‘marker’ animals (with unusual coloration) on paper, and variation in group sizes helps reduce the risk of double counting during aerial surveys. Observers are also to take photographs of many observed groups and use those photos after landing to identify any groups that might have been inadvertently recorded twice. Unfortunately, there is no effective way to correct for the converse problem of burros fleeing and thus never having the opportunity for being detected. Wild burros tend to move more slowly than wild horses. Despite this, because observers can account for burro movements leading to double counting, but cannot account for movement causing burros to never be observed, animal movements can contribute to the estimated abundance (Table 1) potentially being lower than true abundance.

4. The simultaneous double observer method assumes that all burro groups with identical sighting covariate values have equal sighting probability. If there is additional variability in sighting probability not accounted for in the sighting models, such heterogeneity could lead to a negative bias (underestimate) of abundance. In other words, under most conditions the double-observer method underestimates abundance.

5. The analysis assumes that the number of animals in each group is counted accurately. Standard Operating Procedures (Griffin et al. 2020) specify that all groups with more than 20 animals are photographed and photos scrutinized after the flight to correct counts. Smaller groups, particularly ones with poor sighting conditions such as heavy tree cover, could also be undercounted. Any such undercounting would lead to biased estimates of abundance.

### **Evaluation of Survey and Recommendations**

It appears that survey protocols were followed well and with enough consistency among surveys to enable useful pooling of data for more precise estimates of sighting probability. Observers appear to have been well trained, and visibility conditions were largely very good to excellent. However, even the estimated abundance values in Table 1 should be viewed as underestimates of actual wild burro population size in the Black Mountain HMA because of the assumptions and caveats listed above, and because it is known that true burro group detection probabilities tend to

be lower than values estimated by the double-observer data analysis method (Hennig et al., USGS Fort Collins Science Center, manuscript in review).

The survey covered the majority of the HMA and extended beyond the HMA boundaries in a number of places (Figure 1). Areas within the HMA boundary that were not covered were mostly residential areas except for the portion of the HMA that lies north of Hwy 93. According to local BLM and AZGFD personnel, and based on visual observation of burros and burro sign, there are none to very few burros in that area. Where the survey covered areas outside of the HMA, burros were sometimes observed near the edge of the extended survey area. Consequently, it is difficult to be sure there were no additional burros inside or outside of the HMA in areas not surveyed and results should be understood to represent the burros present only in the area surveyed, which may not represent all burros that occasionally occupy this area. Therefore, careful consideration should be given to where burros were located near the edge of the area surveyed when planning whether to extend the survey area further in future surveys to ensure covering all areas potentially occupied by burros associated with the HMA, or to confirm that the current survey boundaries do cover the full extent of burros' range in this area.

**Table 2.** Tally of raw counts of burros and burro groups by observer (front, back, and both) for combined data from Black Mountains HMA surveyed in Nov 2021.

<b>Observer</b>	<b>Groups seen<sup>a</sup> (raw count)</b>	<b>Burros seen (raw count)</b>	<b>Actual sighting rate<sup>b</sup> (groups)</b>	<b>Actual sighting rate<sup>b</sup> (burros)</b>
<b>Front</b>	303	1,658	62.3%	70.0%
<b>Back</b>	362	1,902	74.5%	80.3%
<b>Both</b>	179	1,190	36.8%	50.2%
<b>Combined</b>	486	2,370		

<sup>a</sup> Includes only groups and burros where protocol was followed.

<sup>b</sup> Percentage of all groups seen that were seen by each observer.

**Table 3.** Effect of observers and sighting condition covariates on estimated sighting probability of burro groups for both front and rear observers during the November 2021 survey of Black Mountains HMA. Baseline case (**bold**) for burros presents the predicted sighting probability for a group of 3 burros (the median group size observed) that are not moving, <100 meters from the transect, in 0% vegetation cover, in smooth terrain, not in high contrast lighting, not on the pilot side, with EB as the back-seat observer. Other example cases vary a covariate or observer, one effect at time, as indicated in the left-most column, to illustrate the relative magnitude of each effect. Sighting probabilities for each row should be compared to the baseline (first row) to see the effect of the change in each observer or condition. Baseline values are shown in bold wherever they occur. Sighting probabilities are weighted averages across all 16 models considered (Burnham and Anderson 2002).

	<b>Sighting Probability</b>		
	<b>Front Observer<sup>a</sup></b>	<b>Back Observer<sup>b</sup></b>	<b>Combined Observers</b>
<b>Baseline</b>	64.4%	74.0%	90.7%
<b>Effect of group size (N=1)</b>	58.2%	68.7%	86.9%
<b>Effect of group size (N=10)</b>	81.9%	87.7%	97.8%
<b>Effect of distance = 100-200 m</b>	56.2%	66.9%	85.5%
<b>Effect of distance = 200-300 m</b>	47.6%	58.9%	78.5%
<b>Effect of distance = 300-400 m</b>	39.1%	50.4%	69.8%
<b>Effect of distance = &gt;400 m</b>	21.4%	30.0%	45.0%
<b>Effect of Moving</b>	67.6%	76.7%	92.5%
<b>Effect of Veg2 0.36 (60%)</b>	54.9%	65.2%	84.3%
<b>Effect of Veg2 0.64 (80%)</b>	48.4%	57.9%	78.3%
<b>Effect of Rugged</b>	48.6%	59.8%	79.3%
<b>Effect of Light High</b>	66.0%	75.4%	91.6%
<b>Effect of Pilot Side</b>	29.3%	74.0%	81.6%
<b>Effect of Pilot CJ</b>	88.7%	74.0%	97.1%
<b>Effect of observer MC</b>	64.4%	65.9%	87.9%
<b>Effect of observer CB</b>	64.4%	82.2%	93.7%
<b>Effect of observer LH</b>	64.4%	72.7%	90.3%
<b>Effect of observer CM</b>	64.4%	37.2%	77.6%
<b>Effect of observer BF</b>	64.4%	31.6%	75.6%
<b>Effect of back=front</b>	64.4%	64.4%	87.3%

<sup>a</sup> Sighting probability for the front observers acting as a team, regardless of which of the front observers saw the burros first.

<sup>b</sup> Sighting probabilities for back observers for burro groups that are potentially visible on the same side of the aircraft as the observer. Sighting probability in the back is 0 for groups on the opposite side or centerline.



## Literature Cited

Bureau of Land Management. 2010. Wild horse and burro population inventory and estimation: Bureau of Land Management Instructional Memorandum No. 2010-057. 4 p.

Burnham, K., and D. R. Anderson. 2002. Model Selection and Multimodel Inference: A Practical Information-Theoretic Approach. Springer-Verlag, New York, New York.

Ekernas, L. S., and B. C. Lubow. 2019. R script to analyze wild horse and burro double-observer aerial surveys. USGS Software Release.

Griffin, P. C., L.S. Ekernas, K.A. Schoenecker, and B. C. Lubow. 2020. Standard Operating Procedures for wild horse and burro double-observer aerial surveys. U.S. Geological Survey Techniques and Methods, book 2, chap. A16, 76 p., <https://doi.org/10.3133/tm2A16>.

Lubow, B. C., and J. I. Ransom. 2016. Practical bias correction in aerial surveys of large mammals: validation of hybrid double-observer with sightability method against known abundance of feral horse (*Equus caballus*) populations. PLoS-ONE 11(5):e0154902. doi:10.1371/journal.pone.0154902.

National Research Council. 2013. Using Science to Improve the BLM Wild Horse and Burro Program. The National Academies Press. Washington, D.C.

**Figure 1.** Map of survey tracks flown (black lines), locations of observed burro groups (black and white circles), and HMA boundaries (blue).

